

1 **SALINITY MODULATES THERMOTOLERANCE, ENERGY METABOLISM AND**
2 **STRESS RESPONSE IN AMPHIPODS *GAMMARUS LACUSTRIS***

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4 Vereshchagina K.P.¹, Shatilina Z.M.^{1,2}, Bedulina D.S.¹, Gurkov A.N.¹, Axenov-Gribanov
5 D.V.^{1,2}, Baduev B.K.¹, Kondrateva E.S.¹, Gubanov M.V.³, Zadereev E.S.^{3,4}, Sokolova I.M.⁵ and
6 Timofeyev M.A.¹

7

8 ¹ Irkutsk State University, Karl Marx St. 1, 664003, Irkutsk Russia

9 ² Baikal Research Centre, Lenina 3, 664003, Irkutsk Russia

³ Institute of Biophysics SB RAS, 660036 Krasnoyarsk, Akademgorodok, Russia

⁴ Siberian Federal University, 660041, Svobodnyi 79, Krasnoyarsk, Russia

10 ⁵ Marine Biology, Institute for Biological Sciences, University of Rostock, Rostock, Germany

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13 **ABSTRACT**

14 Temperature and salinity are important abiotic factors for aquatic invertebrates. We
15 investigated the influence of different salinity regimes on thermotolerance, energy metabolism
16 and cellular stress defense mechanisms in amphipods *Gammarus lacustris* Sars from two
17 populations. We exposed amphipods to different thermal scenarios and determined their
18 survival as well as activity of major antioxidant enzymes (peroxidase, catalase, glutathione S-
19 transferase) and parameters of energy metabolism (content of glucose, glycogen, ATP, ADP,
20 AMP and lactate). Amphipods from a freshwater population were more sensitive to the thermal
21 challenge, showing higher mortality during acute and gradual temperature change compared to
22 their counterparts from a saline lake. A more thermotolerant population from a saline lake had
23 high activity of antioxidant enzymes. The energy limitations of the freshwater population
24 (indicated by low baseline glucose levels, downward shift of the critical temperature of aerobic
25 metabolism and inability to maintain steady-state ATP levels during warming) was observed,
26 possibly reflecting a trade-off between the energy demands for osmoregulation under the hypo-
27 osmotic condition of a freshwater environment and protection against temperature stress.

28

30 Temperature and salinity are important abiotic factors affecting survival and
31 performance of aquatic invertebrates and setting limits to their geographical distribution
32 (Kinne, 1971; Nelson, 1977; Pörtner and Farrell, 2008). Recent data indicate the effect of the
33 global climate change to surface waters including lakes worldwide (O'Reilly et al., 2015). The
34 bulk of aquatic organisms are ectothermic, that makes them particularly responsive to one of
35 the main effects of global climate change: the increased variability in temperature, due to the
36 dependency of their body temperature and the metabolic processes from the ambient water
37 temperature. Temperature strongly affects ectotherms due to its direct effects on the rates of
38 physiological and biochemical reactions (Hochachka, 1973; Cossins et al., 1995; Somero,
39 2004). Temperature fluctuations often co-occur with other selective pressures, such as
40 variations in the salt content of the water. In aquatic organisms, deviations from the species-
41 specific optimal salinity can result in osmotic stress requiring adjustments of the cell volume
42 and modulation of enzyme activity to maintain normal cellular functions under the altered ionic
43 conditions; these adjustments can lead to elevated energy demand for osmotic, ionic and/or cell
44 volume regulation (Glazier and Sparks, 1997; Kinne, 1971; Freiere et al. 2011). Therefore,
45 environmental stress (including salinity and temperature stress) can significantly influence the
46 energy balance of living organisms due to the additional energy required to restore and maintain
47 homeostasis, which can put a strain on the energy acquisition, transformation and conservation
48 systems (Sokolova et al, 2012; Helmut Acker et al, 1988; Calow and Forbes, 1998; Amiard-
49 Triquet et al, 2011).

50 Antioxidant systems play a major role in environmental stress response. Antioxidants
51 maintain the cellular redox balance and prevent excess of reactive oxygen species (ROS) from
52 interacting with the critical intracellular structures (Circu and Aw, 2010; Jones, 2008).
53 Antioxidant responses are involved in responses to thermal stress in aquatic organisms (Abele
54 et al., 2007; Sokolova et al., 2011), and earlier studies suggest that some stressors (such as trace
55 metals and hypoxia) that affect cellular redox balance, can modulate the cellular response to
56 thermal stress in aquatic organisms (Ivanina et al., 2009; Pörtner and Langenbuch, 2005;
57 Sokolova and Lannig, 2008). However, the effects of environmental salinity on
58 thermotolerance and cellular protection mechanisms of aquatic organisms are not well
59 understood and require further investigation.

60 Freshwater amphipods of the genus *Gammarus* Fabricius, 1775 are a useful model to
61 study the interactive effects of salinity and thermal stress. *Gammarus* is a widespread genus of

62 amphipods in the northern hemisphere that plays a key role in the food webs of freshwater
63 ecosystems. Some freshwater amphipods (including those of genus *Gammarus*) occur in
64 habitats ranging from low mineralization fresh waters to brackish waters. These species also
65 experience wide diurnal and seasonal temperature fluctuations in temperate shallow waters
66 (Jacobs et al, 1997) and thus must possess efficient mechanisms of protection against
67 temperature-induced cellular stress. The molecular and cellular mechanisms of the broad
68 salinity and temperature tolerance of freshwater amphipods remain to be fully elucidated
69 (Grzesiuk and Mikulski, 2006).

70 The aim of this study was to determine whether acclimation/adaptation to habitats with
71 different salinity regimes modulates the energy metabolism, cellular protective responses to
72 temperature stress and thermal tolerance of amphipod *Gammarus lacustris* Sars (Sars, 1863).
73 We exposed amphipods from two populations adapted to different salinity regimes to several
74 thermal scenarios (namely hypothermia, as well as acute and gradual warming) and determined
75 their survival, activity of major antioxidant enzymes (peroxidase, catalase, glutathione S-
76 transferase) and parameters of energetic metabolism (content of glucose, glycogen, ATP, ADP,
77 AMP and lactate) in order to gain insights into the physiological and cellular mechanisms of
78 temperature-salinity interactions in this ecologically important eurybiont species.

79

80 **2. MATERIALS AND METHODS**

81 **2.1 Animals**

82 Holarctic amphipod *Gammarus lacustris* Sars, 1863 is a widespread omnivore species that
83 inhabits lentic and lotic ecosystems and has a broad tolerance to environmental stressors
84 (Karaman, Pinkster, 1977; Barnard, Barnard, 1983; Matafonov, 2007; Väinölä, 2007; Takhteev,
85 2015). This species reproduces during the summer and can have several reproduction periods
86 depending on the environmental and ecological characteristics of the water bodies (Timoshkin,
87 2001, 2008). Under the experimental conditions the preferred temperatures for *G. lacustris* is
88 15 - 16°C (Timofeyev, 2010). It is a euryhaline species with a broad pH tolerance (6.2 to 9.2)
89 that inhabits benthic and pelagic zones of the lakes and is often the top predator in the absence
90 of fish (Wilhelm and Schindler, 1999; Matafonov, 2007).

91 **2.2 Sampling locations**

92 We collected amphipods in July of 2009, 2013 and 2014 from two different habitats (a
93 freshwater habitat and a saline lake) within Eastern and Western Siberia (Russia). A freshwater

94 population of *G. lacustris* was collected from a shallow lake formed by a backwater of Angara
 95 River in the urban area of Irkutsk city (52°16'4.71" N, 104°16'52.77" E). Amphipods were
 96 sampled by a hand net from the depth 0-1 m. Water samples were collected at the same time
 97 and analyzed for ionic composition by the Interinstitutional Regional Laboratory of
 98 Environmental Research at Irkutsk State University. Specimens of amphipods from a saltwater
 99 population were obtained from the southern shore of a meromictic lake Shira (54°29'7.25" N,
 100 90°12'1.49" E) from the depth of 7 m using a plankton net. The freshwater site has pH 8.4 and
 101 low mineralization (0.5 g L⁻¹) of the water (Table 1). The saline lake has a high content of
 102 dissolved minerals (15–17 g l⁻¹) with Na⁺, Mg²⁺, and Ca²⁺ as the major cations (Kalacheva et
 103 al., 2002) (Table 1). Thermal regimes are similar at the two study sites. Both lakes can
 104 completely freeze in winter, and summer temperature may reach 23°C in the near-shore waters.
 105 Annual average temperature in both waterbodies is 6-7°C (Rogozin et al. 2010).

106 Table 1. Chemical composition of the surface water from brackish Lake Shira and a freshwater
 107 lake in Irkutsk (Data for Lake Shira from Kalacheva et al., 2002)

Lake	Ions	Cl ⁻	Na ⁺	K ⁺	Mg ²⁺ ₊	CO ₃ ²⁻	Ca ²⁺	SO ₄ ²⁻	HCO ₃ ⁻	pH	Mineralization g/L
Shira	mg/ L	2100	2880	37	1080	174	51	8010	998	8.7	16.60
In Irkutsk		55.8	47.6	1.6	35.7	15	48.1	105	223	8.4	0.53

108 2.3 Experimental design

109 Two types of experiments were carried out in this study: an acute temperature stress and
 110 gradual warming. Prior to the exposures animals were pre-acclimated for 3 – 7 days under the
 111 constant aeration in filtered water from their native habitats. Pre-acclimation was conducted at
 112 the temperature recorded at the time of sampling (15°C for both populations) or at 7°C (as
 113 annual average temperature of both waterbodies) in the case of the gradual warming.
 114 Amphipods were fed daily with natural food (elm leaves) with addition of commercial food
 115 (Tetra-Min, Tetra GmbH, Germany) and potato *ad libitum*. No mortality was observed during
 116 pre-acclimation. Only actively swimming animals were used for experiments. Control animals
 117 for all experiments were kept under the same conditions as during the preliminary acclimation.

118 *Mortality*

119 To determine median lethal times (LT50) that cause 50% mortality during acute heat
 120 exposure, 10 individuals from each of the two studied populations were placed in separate tanks
 121 filled with 1 L of the filtered water from their respective habitats pre-heated to 30 °C. Nine

122 replicate tanks were used for each of the studied populations, to the total of 90 animals per
123 population. Mortality was monitored every hour until all animals died.

124 *Acute thermal stress*

125 Amphipods were placed in 2.5 L aerated glass tanks (n = 5) with water, pre-heated to
126 30 °C (according to our observations, the maximum temperature in the littoral zones of lake in
127 Irkutsk can reach up to 30 °C in a scorching summer). After 0.5, 1, 3 and 6 hours of exposure,
128 3 individuals from each of the five replicate tanks were collected and flash-frozen in liquid
129 nitrogen for subsequent biochemical analyses.

130 *Gradual temperature increase*

131 Animals were pre-acclimated for 7 days at the temperature of 7 °C in separate glass
132 tanks (2.5 L, n = 7) and subjected to the gradual temperature increase of 1 °C per hour
133 continuing until 100% mortality occurred (modified from Sokolova and Pörtner (2003)). After
134 every 2 degrees of temperature increase (i.e. every 2 hours), 3 specimens were collected from
135 each tank of the seven replicate tanks and shock-frozen in liquid nitrogen.

136 **2.4 Biochemical methods**

137 *Enzymatic activities*

138 Activities of total cellular peroxidases, as well as catalase and glutathione S-transferase
139 were measured using standard spectrophotometric assays. Enzyme extraction was done as
140 described elsewhere (Bedulina et al., 2010) using a 1:3 (w:v) ratio of homogenization medium
141 to amphipod biomass. Enzyme activities were measured at 25°C in the supernatant using a
142 SmartSpec Plus spectrophotometer (Bio-Rad, Hercules, USA) and Cary 50 spectrophotometer
143 (Varian, USA). Total peroxidase activity in the soluble fraction was measured with 4.4 mM
144 guaiacol as a substrate at 436 nm, pH 5, according to Bergmeyer (1974). Catalase activity was
145 measured with 2.25 mM hydrogen peroxide as a substrate at 240 nm, pH 7, according to Aebi
146 (1984). Glutathione S-transferase activity was measured with 0.97 mM 1-chloro-2,4-
147 dinitrobenzene (CDNB) as a substrate at 340 nm, pH 6.5 (Habig et al., 1974). Bradford assay
148 was used to evaluate protein concentrations (Bradford, 1976). Enzyme activities were expressed
149 in nKatal mg⁻¹ protein. For each enzyme and experimental condition, we measured 4-9
150 biological replicates, each replicate consisting of the pooled tissues of two individuals.

151 ***Lactate content***

152 Lactate levels were determined using the express-kit “Lactate-vital” (Vital–
153 Diagnostics, St. Petersburg, Russia) according to Bergmeyer (1985). Absorbance was measured
154 with a Cary 50 spectrophotometer (Varian, USA) at $\lambda=505$ nm.

155 ***Metabolites***

156 The levels of glucose, glycogen, ATP, ADP and AMP were measured using the
157 NADH/NADPH-dependent enzymatic methods at $\lambda = 340$ nm according to Grieshaber et al.
158 (1994), Bergmeyer (1985) and Sokolova et al. (2012). Glycogen was hydrolyzed using the
159 methods of Sokolova et al. (2012). Absorbance was measured with a Cary 50
160 spectrophotometer (Varian, USA).

161 The adenylate energy charge was calculated according to Atkinson, 1968, using the
162 following equation $AEC = (ATP+0.5ADP)/(ATP+ADP+AMP)$.

163 **2.5 Statistical analyses**

164 Mortality rates in both populations under heat shock conditions were fitted to the
165 Weibull model (Wilson, 1994) in statistical package R (R Core Team, 2016), and LT50 values
166 (times at which mortality of 50% individuals occurred) were derived from the build regression
167 model:

168
$$m = 100 - \frac{100}{e^{\left(\frac{t}{p}\right)^r}}$$

169 m – cumulative mortality, %

170 t – time, h

171 p and r – regression coefficients.

172

173 All experiments were carried out with 3-8 biological replicates, and biochemical
174 measurements were performed in triplicate (technical replicates) for each sample. Normality
175 was tested by Kolmogorov-Smirnov test and the equal variance with Levene’s test. Data were
176 analyzed statistically using one-way analysis of variance (ANOVA, general linear model),
177 followed by Student–Newman–Keuls post hoc test. If the date distribution was not normal,
178 Kruskal-Wallis with Dunn’s test was applied. Differences were considered significant at p -
179 value < 0.05 (after corrections for multiple comparisons). Statistical analysis was carried out
180 using the software package SigmaPlot (version 12, Systat Software Inc., USA/Canada).

181

182 **RESULTS**

183 **Mortality**

184 Median mortality times (LT50) during acute exposure to 30°C were significantly higher in
185 amphipods from the saline Lake Shira (LT50 = 22.8 h) compared to their counterparts from a
186 freshwater Lake in Irkutsk (LT50 = 7.7 h) (Fig. 1). 100% mortality during gradual warming
187 was observed at 31°C in Irkutsk population and 33°C in Shira population (data not shown).

188

189 **Activity of antioxidant enzymes**

190 **Peroxidase**

191 *Effect of acclimation at different temperatures*

192 The levels of peroxidase activity were significantly higher after the acclimation at 7°C
193 than at 15°C in amphipods from both populations (Fig. 2; A).

194 *Exposure to heat shock challenge*

195 The basal levels of peroxidase activity at 15°C were similar in amphipods from the
196 freshwater and saltwater habitats. However, in amphipods from the freshwater Irkutsk site
197 peroxidase activity increased after 1 h of heat shock exposure (30°C) and remained elevated
198 until the end of the heat exposure period. In amphipods from the saltwater site, total peroxidase
199 activity was not significantly affected by the heat shock exposure. As a result, amphipods from
200 a freshwater Irkutsk habitat had a higher peroxidase activity during 1-3 h of heat shock exposure
201 compared to their saltwater counterparts (Fig. 3A).

202 *Exposure to the gradual thermal challenge*

203 The level of peroxidase activity at 7°C was significantly higher in amphipods from the
204 saltwater site compared to their freshwater counterparts. Beyond the exposure time, peroxidase
205 activities changed significantly in amphipods from both populations. In the freshwater
206 population of amphipods, peroxidase activity remained at the steady-state levels in the range of
207 7 - 17°C and significantly decreased at 19 °C. In amphipods from Shira population, gradual
208 warming resulted in a significant decrease of peroxidase activity at 11°C, and peroxidase
209 activity remained suppressed throughout the rest of the thermal exposure period (Fig. 3B).

210

211 **Glutathione S-transferase (GST)**

212 *Effect of acclimation at different temperatures*

213 The level of GST activity was lower in amphipods from the freshwater population,
214 acclimated at 7°C compared to 15°C, however in amphipods from the saltwater population the
215 activity of this enzyme was significantly higher after the acclimation at 7°C than at 15°C (Fig.
216 2, B).

217 *Exposure to heat shock challenge*

218 The levels of GST activity at 15°C were similar in amphipods from the freshwater and
219 saltwater habitats. In amphipods from the freshwater Irkutsk site, there was significant increase
220 in GST activity after 1 and 3h of heat exposure. In amphipods from the saltwater site, heat
221 exposure led to a transient decline in GST activity after 0.5 - 1 h of heat exposure, which
222 returned the basal levels after 3-6 h of the heat shock (Fig. 3C).

223 *Exposure to a gradual thermal challenge*

224 The baseline levels of GST activity at 7°C were significantly higher in amphipods from
225 the saltwater population compared to the freshwater population of amphipods. In both
226 populations of amphipods the activity of GST was constant throughout the exposure to gradual
227 warming (Fig. 3D).

228 **Catalase**

229 *Effect of acclimation at different temperatures*

230 The level of catalase activity was significantly higher in amphipods, acclimated at 7°C
231 than at 15°C, from the both studied populations (Fig. 2, C).

232

233 *Exposure to heat shock challenge*

234 The baseline levels of catalase activity at 15°C were significantly higher in amphipods
235 from the saltwater site compared to their freshwater counterparts. Catalase activity significantly
236 increased in amphipods from the freshwater habitat after 1 and 3h of heat exposure but
237 decreased in their counterparts from the saltwater site after 3h in response to the acute heat
238 exposure (Fig. 4A).

239

240 *Exposure to a gradual thermal challenge*

241 The baseline levels of catalase activity at 7°C were significantly higher in amphipods
242 from the saltwater site compared to their freshwater counterparts. Gradual temperature increase
243 did not affect the activity of catalase in both of the studied populations of amphipods throughout
244 the entire range of experimental temperatures (Fig. 4B).

245 **Lactate content**

246 *Effect of acclimation at different temperatures*

247 The level of lactate was significantly lower in amphipods from the freshwater
248 population, acclimated at 7°C, compared to 15°C. Acclimation at different temperatures had
249 no significant effect on the lactate level in amphipods from the saltwater population (Fig. 2, F).

250 *Exposure to heat shock challenge*

251 The basal levels of lactate at 15°C were significantly higher in amphipods from the
252 freshwater population (Irkutsk) compared to saltwater Shira population. Acute thermal stress
253 resulted in a significant increase of lactate level after 0.5 h and was increased until the end of
254 the exposure (6h) in amphipods from the saline like, whereas content of lactate significantly
255 increased only after 3h of acute thermal stress in amphipods from freshwater site. (Fig. 4C).

256 *Exposure to a gradual thermal challenge*

257 The basal levels of lactate at 7°C were similar in amphipods from both studied
258 populations. In amphipods from the freshwater Irkutsk population, gradual warming resulted in
259 a significant elevation of lactate content at 17°C that continued increasing until the end of the
260 thermal exposure. In amphipods from Shira population, the content of lactate was constant
261 within the temperature range of 7 - 29°C and increased at 31°C and 33 °C. (Fig. 4D).

262

263 **Glucose content**

264 *Effect of acclimation at different temperatures*

265 Acclimation at different temperatures had no significant effect on the glucose level in
266 amphipods from the freshwater population. The level of glucose was significantly lower in
267 amphipods from the freshwater population, acclimated at 7°C, compared to 15°C. (Fig. 2, D).

268 *Exposure to acute thermal challenge*

269 The basal level of glucose at 15°C was 5-fold higher in amphipods from the saltwater
270 site compared to their freshwater counterparts. In amphipods from the freshwater site, acute
271 heat stress led to a significant increase in glucose content after 3 h. In contrast, acute thermal
272 challenge resulted in a significant decrease of glucose content after 3 and 6 hours in amphipods
273 from a saline population (Fig. 5A).

274 *Exposure to a gradual thermal challenge*

275 The basal level of glucose at 7°C was higher in amphipods from the saltwater site
276 compared to their freshwater counterparts (Fig. 5B). In amphipods from the freshwater
277 population, glucose content was at the steady-state in the temperature range of 7 - 27°C and
278 significantly increased at 29 °C. In amphipods from a saline lake, gradual warming resulted in
279 a significant decline of glucose content at 13, 19, 23-29 and 33°C (Fig. 5B).

280

281 **Glycogen content**

282 *Effect of acclimation at different temperatures*

283 The level of glycogen was significantly lower in amphipods from the freshwater
284 population, acclimated at 7°C, compared to 15°C. Opposite to that, the level of glycogen was
285 higher in amphipods from the saltwater population, acclimated at 7°C compared to 15°C (Fig.
286 2, E).

287 *Exposure to heat shock challenge*

288 The basal glycogen content at 15°C was higher in amphipods from the freshwater
289 population (Irkutsk) compared to a saltwater population (Lake Shira) (Fig. 5C). Acute heat
290 stress resulted in a significant decrease of glycogen content in amphipods from the saltwater
291 (after 0.5 and 3 h) and freshwater (after 3 h of exposure) populations.

292 *Exposure to a gradual thermal challenge*

293 At 7°C the basal levels of glycogen were significantly higher in amphipods from the
294 saltwater site compared to their freshwater counterparts. Glycogen content did not change in
295 freshwater or saltwater amphipods during the gradual temperature increase (Fig. 5, D).

296 **Adenylates**

297 *Effect of acclimation at different temperatures*

298 Acclimation at 7 and 15°C had no significant effect on the levels of ATP, ADP and
299 AMP in amphipods from the freshwater population; however, the level of ATP was
300 significantly lower in amphipods from the saltwater population, acclimated at 7°C, compared
301 to 15°C. (Fig. 2; G, H, I).

302 *Exposure to heat shock challenge*

303 The baseline of ATP content at 15°C was significantly higher in amphipods from
304 freshwater site compared to their saltwater counterparts (Fig. 6, A). ATP level did not change
305 in freshwater or saltwater amphipods throughout the gradual temperature increase. The basic
306 level of AMP content at 15 °C was higher in amphipods from the saltwater Shira population
307 compared to the freshwater Irkutsk population. AMP levels remained at the steady-state during
308 acute heat stress in amphipods from a saline lake but increased after 1-3 h of exposure to 30 °C
309 in amphipods from the freshwater site (Fig. 6, C). Baseline ADP levels were similar in
310 amphipods from both studied populations (Fig. 7, A). There were no changes in the content of
311 ADP in amphipods during acute heat stress.

312 *Exposure to a gradual thermal challenge*

313 The baseline ATP content at 7°C was significantly higher in amphipods from the
314 freshwater population compared to the saltwater population. In amphipods from the freshwater
315 site, gradual warming led to a decrease in tissue levels of ATP at the temperature of 9°C or
316 higher. In amphipods from the saltwater Shira site, gradual warming led to a steady increase in
317 tissue ATP levels at the temperature of 13°C or higher (Fig. 6, B). In contrast to ATP, the
318 baseline AMP content was higher in individuals from the saltwater population, compared to
319 their freshwater counterparts. AMP level of freshwater population of amphipods was decreased
320 at 31°C in response to a gradual warming, but there was no changes in AMP content of Shira
321 population (Fig. 6, D). The baseline ADP content was similar in amphipods from the two studied
322 sites. There were no changes in the ADP content in amphipods from both populations (Fig. 7,
323 B).

324 **Adenylate energy charge**

325 *Effect of acclimation at different temperatures*

326 Acclimation at 7 and 15°C had no significant effect on adenylate energy charge levels
327 (AEC) in amphipods from the freshwater population; however, the value of AEC significantly

328 decreased in amphipods from the saltwater population, acclimated at 7°C, compared to 15°C.
329 (Fig. 7; C, D).

330 *Exposure to heat shock challenge*

331 The baseline level of AEC at 15°C was significantly higher in amphipods from the freshwater
332 population compared to the saltwater population. Heat shock exposure led to a decrease of AEC
333 in amphipods from the saltwater population already after 0.5 and 1 hours, whereas no significant
334 change in AEC of the freshwater population was observed (Fig. 7, C).

335 *Exposure to a gradual thermal challenge*

336 The baseline level of AEC after the acclimation at 7°C was significantly higher in amphipods
337 from the freshwater population. Gradual warming led to a significant decrease of AEC in
338 amphipods of the freshwater population starting at 17°C with a short leveling in the temperature
339 range 21 – 25°C and a following decrease at 27 and 29°C. In contrast, the AEC in amphipods
340 from saltwater population significantly increased during the gradual warming, starting from
341 9°C, and the level of AEC stayed elevated during the whole exposure (Fig. 7, D).

342 **Discussion**

343 Our study shows that adaptation to different salinity regimes can influence
344 thermotolerance and modulate key characteristics of cellular metabolism and stress responses
345 in Holarctic amphipods *G. lacustris* Sars. Amphipods from a freshwater Irkutsk population
346 were more sensitive to the thermal challenge experiencing higher mortality during acute and
347 gradual warming compared to the amphipods from the saline Lake Shira. *G. lacustris* is well
348 known for its broad salinity tolerance compared to its congeners such as *G. pulex* (Sutcliffe and
349 Shaw, 1967). Earlier studies showed that freshwater *G. lacustris* can survive in 50 and 60% sea
350 water for at least 8 weeks, and regulate the hemolymph Na⁺ concentration until the media Na⁺
351 content reaches 250 mM L⁻¹ (Sutcliffe and Shaw, 1967). Na⁺ concentration in the freshwater
352 lake in Irkutsk (2.07 mM L⁻¹) is significantly hypo-ionic with respect to *G. lacustris* hemolymph
353 (that contains ~150-200 mM Na⁺ under freshwater and brackish conditions, Sutcliffe and
354 Shaw, 1967), whereas the water in the saline Lake Shira is almost isoionic ([Na⁺]=126 mM L⁻¹)
355 to *G. lacustris* hemolymph. This indicates that *G. lacustris* from the saline lake are less
356 osmotically stressed than their freshwater counterparts and therefore spend less energy on ion-
357 and osmoregulation which may explain their higher thermal tolerance compared to the
358 freshwater population.

359 *Basal levels of cellular stress markers and energy metabolism*

360 Bioenergetic indices (tissue levels of adenylates) measured under the temperature
361 conditions similar to the mean habitat temperature and close to the physiological optimum for
362 *G. lacustris* (15°C), (Timofeev, 2010), indicate more active aerobic metabolism in the
363 amphipods from the freshwater Irkutsk population compared to their saltwater counterparts.
364 Additionally, low levels of lactate in tissues of the saltwater amphipods suggest low rates of
365 anaerobic metabolism (Pörtner et al., 1984; Livingstone, 1983). The freshwater amphipods
366 appear to more strongly depend on anaerobic glycolysis compared to their saltwater
367 counterparts as indicated by low glucose concentration and high levels of lactate, a major end
368 product of anaerobic glycolysis in freshwater crustaceans (Pörtner et al., 1984; Livingstone,
369 1983).

370 Acclimation at a lower temperature (7°C) resulted in a significant stimulation of activity
371 of antioxidants in both saltwater and freshwater amphipod populations. This indicates that
372 hypothermia induces oxidative stress in the amphipods, possibly reflecting higher oxygen
373 solubility at low temperatures. Notably, the degree of stimulation of antioxidants was stronger
374 in the saltwater Shira amphipods. Elevated activities of antioxidant enzymes and/or total
375 antioxidant capacity, were also earlier reported from other euryhaline crustaceans acclimated
376 or acclimatized to high salinities although profiles of antioxidant response to elevated salinity
377 were tissue-specific (Henry et al., 2012; Wang et al, 2013; Romano and Zeng, 2013; Freire et
378 al, 2011; Li et al, 2015).

379 Energy metabolism of amphipods was also affected by the hypothermic conditions
380 (7°C). In saltwater Shira amphipods, the hypothermic acclimation (compared to the optimal
381 temperature of 15°C) led to a decrease in ATP, a slight decline in free glucose and accumulation
382 of glycogen. This indicates stimulation of glycogen synthesis at low temperatures in the
383 saltwater amphipods. In the freshwater population hypothermic exposure had opposite effects
384 on energy metabolites leading to an increase in free glucose, decline in the glycogen reserves
385 while the levels of adenylates remained unchanged. This suggests an increase in the glycolytic
386 ATP production at 7°C and may account for the maintenance of the steady-state ATP levels in
387 hypothermia in the freshwater amphipods.

388

389 *Hyperthermic stress*

390 Acute thermal challenge (30°C) led to a significant increase in activities of all tested
391 antioxidant enzymes (peroxidase, GST and catalase) in the freshwater population of amphipods.
392 Elevated levels of antioxidant enzymes during the thermal challenge may reflect a temperature-
393 induced increase in generation of reactive oxygen species (ROS) in the less tolerant freshwater
394 population requiring upregulation of the cellular antioxidant capacity to protect to organism
395 against oxidative stress. No such increase in antioxidant levels was found in the more
396 thermotolerant saltwater population of amphipods. This may suggest that the existing capacity
397 of antioxidant enzymes is sufficient to protect against the temperature-induced generation of
398 ROS in these organisms and/or that their mitochondria do not increase the rate of ROS release
399 during the heat exposure. Moreover, activity of antioxidant enzymes showed a transient but
400 significant decline in response to the thermal challenge in *G. lacustris* from the saltwater site.
401 This is unlikely to indicate a direct thermal damage to the antioxidant enzymes given high
402 tolerance to heating in this population but may rather reflect a decrease in the ROS production
403 at intermediate temperatures.

404 Gradual temperature increase starting from 7°C resulted in a decrease of peroxidase
405 activity in Shira population at 11-21°C followed by a slight increase at 27-33°C, which,
406 however, did not reach the control levels. In freshwater Irkutsk population, peroxidase activity
407 declined at 19-21°C then returned to the control levels. It is worth noting that the total
408 peroxidase activity measured in this study encompasses activities of diverse cellular
409 peroxidases such as peroxiredoxins, glutathione peroxidases, peroxisomal peroxidases and
410 others. These enzymes are involved in a broad variety of cellular functions but they all play a
411 role in the cellular redox balance and control of the oxidative stress (Mates, 2000; Espinosa-
412 Diez et al., 2015; García-Triana, 2012). A decrease in peroxidase activity at intermediate
413 temperatures ($\pm 6^\circ\text{C}$ around the optimum) may reflect either a decrease in ROS production, or
414 serve as an energy-saving mechanism that redirects resources to other protective mechanisms
415 (e.g. chaperones) to ensure optimal cellular protection at high temperatures. A similar decrease
416 in activities of antioxidant enzymes has been previously shown in response to a variety of
417 environmental stressors in amphipods (Timofeyev et al., 2008; Timofeyev et al., 2009). No
418 changes in activity of catalase and glutathione S-transferase were found in the freshwater or
419 saltwater populations of amphipods. Taken together, these findings indicate that baseline levels
420 of antioxidant enzymes are sufficient to cope with the temperature-induced oxidative stress
421 during gradual warming, and/or that unlike the acute thermal stress, gradual warming does not
422 induce elevated ROS production in *G. lacustris*.

423 It is worth noting that our present study, while focusing on three key antioxidant
424 enzymes, did not exhaustively test all cellular antioxidants that can be involved in protective
425 responses against temperature and salinity challenge in aquatic organisms (Mates, 2000;
426 Espinosa-Diez et al., 2015; García-Triana, 2012). Lower baseline levels of some antioxidants
427 (such as catalase) and a lack of increase in antioxidant levels during heat stress in the saltwater
428 amphipods may reflect the reliance of these organisms on other antioxidants, such as low
429 molecular weight antioxidants obtained from the diet (Lesser, 2006). An earlier study showed
430 that dietary antioxidants can strongly affect the degree and direction of the antioxidant response
431 to environmental stressors in amphipods (Timofeyev et al., 2008). The differences in the
432 predominant diet in amphipods from the saltwater and freshwater habitats could affect the tissue
433 levels of non-enzymatic antioxidants. Thus, the main food source for the Lake Shira benthic
434 planktonic population of *G. lacustris* are freshly sedimented seston (Gladyshev et al. 2002,
435 Degermendzhy et al. 2010) and planktonic copepods (Yemelyanova et al., 2002) which are rich
436 in low molecular weight antioxidants. In contrast, amphipods from the freshwater Irkutsk lake
437 are benthic and primarily feed on detritus which is low in antioxidant content. This difference
438 in diets may influence the oxidative stress response to heating in the two studied populations
439 and requires further investigation. Regardless of the contribution of the low molecular weight
440 antioxidants to the total antioxidant capacity and stress tolerance of *G. lacustris*, our data
441 demonstrate that activation of enzymatic antioxidants is not involved in elevated heat tolerance
442 of the saltwater amphipod population.

443 Metabolic responses to thermal challenge notably differed in the freshwater and
444 saltwater populations of the studied amphipods. Significantly higher baseline levels of glucose
445 in the saltwater population serves to provide energy resources during the acute thermal
446 challenge (30°C) until the 3h of exposure when the level of glucose dropped about 5-fold. Rapid
447 accumulation of lactate indicates activation of anaerobic glycolysis as early as after 0.5 h of
448 heat exposure in the saltwater population; this would provide additional ATP as the
449 temperature-induced ATP demand of the tissues outstrips the aerobic ATP supply (Mathews et
450 al., 2000; Larade and Storey, 2002; Philp et al., 2005). The stable levels of ATP during the
451 acute temperature stress in the saltwater amphipods indicates that the compensatory onset of
452 anaerobic ATP production in combination with the aerobically produced ATP is sufficient to
453 prevent ATP depletion. Significantly lower baseline levels of glucose in a freshwater population
454 of amphipods indicates that this population may be energy-limited. Activation of
455 glycogenolysis (as indicated by glycogen depletion and up to 5-fold glucose accumulation), and
456 the onset of anaerobiosis (indicated by lactate accumulation) occurs much later in the freshwater

457 amphipods (after 3 h of exposure), which is close to the LT100 for this population. Such late
458 transition to anaerobic metabolism and glycogenolysis indicate limited metabolic plasticity of
459 the freshwater population during heat stress and may contribute to their higher sensitivity to
460 temperature (and potentially other stressors).

461 Induction of anaerobiosis (as indicated by lactate accumulation) during the gradual
462 warming occurred at considerably lower temperatures in the freshwater amphipods (17°C)
463 compared to their saltwater counterparts (31°C). This indicates that the temperature-induced
464 mismatch between the ATP demand and aerobic ATP supply occurs at lower temperatures in
465 the less thermotolerant freshwater amphipods. Given the ability of saltwater amphipods to
466 rapidly engage anaerobic pathways during acute thermal stress (30°C), the delayed onset of
467 anaerobiosis during the gradual temperature rise indicates that aerobic energy supply is
468 sufficient to support ATP turnover until very high temperatures (31°C) are reached. Transition
469 to partial anaerobiosis during gradual warming is considered an indication of the upper critical
470 temperatures of aerobic metabolism (T_{critII}) as which the aerobic scope of the organism
471 disappears and only time-limited survival is possible (Sokolova et al., 2012, Sokolova &
472 Bagwe, 2015). The upper critical temperatures of aerobic metabolism (17°C and 31°C) are
473 lower than the upper thermal limits (31°C and 33°C in the freshwater and the saltwater
474 populations, respectively) and are considered a better indicator of the ecologically relevant
475 thermal limits that determine biogeographic distribution of a species (Pörtner & Knust, 2007,
476 Deutsch et al., 2015).

477 Even though anaerobic glycolysis is not engaged until 31°C in the saltwater population
478 of amphipods, a strong decrease in glucose content in the absence of glycogen accumulation at
479 the temperatures at or above 15°C indicate elevated catabolism of glucose. The most likely
480 pathway for this catabolism is aerobic oxidation of glucose, which may explain an increase in
481 ATP concentrations in saltwater amphipods during gradual warming. This indicates that the
482 saltwater population maintains high cellular energy status over the most environmentally
483 relevant temperature range. In contrast, the freshwater population of amphipods experiences
484 energy deficiency during gradual warming indicated by a decline in ATP content and (at the
485 extreme temperature of 31°C) a decrease in AMP levels suggesting transamination of AMP.

486 No significant change in glycogen levels and only a slight increase of glucose level at
487 29°C is observed during gradual warming in the freshwater population of amphipods. This
488 finding indicates that glucose used for lactate production at elevated temperatures must be
489 replenished by gluconeogenesis from other sources besides the glycogen (such as amino acids)
490 in the freshwater amphipods. Generically, these findings indicate that a decrease in the aerobic

491 scope during the gradual warming in the freshwater amphipods goes hand-in-hand with the
492 impairment of the cellular energy status indicated by the decrease in ATP levels..

493 Overall, our data indicate that higher thermal sensitivity of the freshwater population of
494 amphipods is associated with a lower baseline activity of antioxidant enzymes and a decreased
495 ability to maintain energy balance and curb oxidative stress, compared to their saltwater
496 counterparts, during exposure to acute and gradual temperature increase. High sensitivity of the
497 freshwater population to warming was associated with energy limitations (indicated by low
498 baseline glucose levels, downward shift of the critical temperature of aerobic metabolism and
499 inability to maintain the steady-state ATP levels during warming), possibly reflecting a trade-
500 off between the energy demands for osmoregulation and protection against the temperature
501 stress. These findings suggest that freshwater populations of amphipods may be more
502 vulnerable to the global climate change than those from saline habitats. On the other hand,
503 brackish waters may serve as potential refuges during the climate change for eurysaline
504 amphipod species such as *G. lacustris*.

505

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511

512 **References**

513 Abele D, Philipp E, Gonzalez PM, Puntarulo S. 2007. Marine invertebrate mitochondria
514 and oxidative stress. *Frontiers in Bioscience* 12:933–946. hdl:10013/epic.24770

515 Acker H. 1988. *Oxygen sensing in tissues*. Springer-Verlag.

516 Amiard-Triquet C, Rainbow PS, Roméo M. 2011. *Tolerance to environmental*
517 *contaminants*. CRC Press.

518 Atkinson DE. 1968. Energy charge of the adenylate pool as a regulatory parameter.
519 Interaction with feedback modifiers. *Biochemistry* 7(11):4030-4034.

520 Barnard JL, Barnard CM. 1983. *Freshwater amphipods of the World. I. Evolutionary*
521 *patterns. II. Handbook and bibliography*. Hayfield Associates: Mt. Vernon, Virginia.

- 522 Bedulina DS, Zimmer M, Timofeyev MA. 2010. Sub-littoral and supra-littoral amphipods
523 respond differently to acute thermal stress. *Comparative Biochemistry and Physiology -*
524 *Part B: Biochemistry & Molecular Biology* 155:413–418.
- 525 Bergmeyer HU. 1985. Methods of enzymatic analysis. *VCH Weinheim* 86:648–653.
- 526 Bradford MM. 1976. A rapid and sensitive method for the quantitation of microgram
527 quantities of protein utilizing the principle of protein-dye binding. *Analytical Biochemistry*
528 72:248–254.
- 529 Calow P, Forbes VE. 1998. How do physiological responses to stress translate into
530 ecological and evolutionary processes? *Comparative Biochemistry and Physiology - Part*
531 *A: Molecular & Integrative Physiology* 120(1):11 – 16.
- 532 Circu ML, Aw TY. 2010. Reactive oxygen species, cellular redox systems, and apoptosis.
533 *Free Radical Biology & Medicine* 48(6):749 – 762.
- 534 Cossins AR, Schwarzbaum PJ, Wieser W. 1995. Chapter 6 Effects of temperature on
535 cellular ion regulation and membrane transport systems. *Biochemistry and Molecular*
536 *Biology of Fishes* 5:101–12.
- 537 Degermendzhy AG, Zadereev ES, Rogozin DY, Prokopkin IG, Barkhatov YV, Tolomeev
538 AP, Gulati RD. 2010. Vertical stratification of physical, chemical and biological
539 components in two saline lakes Shira and Shunet (South Siberia, Russia). *Aquatic Ecology*
540 44(3):619 – 632.
- 541 Espinosa-Diez C, Miguel V, Mennerich D, Kietzmann T, Sánchez-Pérez P, Cadenas S,
542 Lamas S. 2015 Antioxidant responses and cellular adjustments to oxidative stress. *Redox*
543 *biology* 6:183 – 197.
- 544 Freire CA, Togni VG, Hermes-Lima M. 2011. Responses of free radical metabolism to air
545 exposure or salinity stress, in crabs (*Callinectes danae* and *C. ornatus*) with different
546 estuarine distributions. *Comparative Biochemistry and Physiology - Part A: Molecular &*
547 *Integrative Physiology* 160(2):291–300.

- 548 Freire CA, Welker AF, Storey JM, Storey KB, Hermes-Lima M. 2011. Oxidative stress in
549 estuarine and intertidal environments (temperate and tropical). *Oxidative stress in aquatic*
550 *ecosystems* 41-57.
- 551 García-Triana A, Yepiz-Plascencia G. 2012. The crustacean selenoproteome similarity to
552 other arthropods homologs: A mini review. *Electronic Journal of Biotechnology* 15(5):17–
553 17.
- 554 Gladyshev MI, Emelianova AY, Kalachova GS, Zotina TA, Gaevsky NA, Zhilenkov MD.
555 2000. Gut content analysis of *Gammarus lacustris* from a Siberian lake using biochemical
556 and biophysical methods. *Hydrobiologia* 43:155–163.
- 557 Glazier DS, Sparks BL. 1997. Forum. *Functional Ecology* 11(1):126-128.
- 558 Grieshaber MK, Hardewig I, Kreutzer U, Pörtner HO. 1994. Physiological and metabolic
559 responses to hypoxia in invertebrates. *Reviews of Physiology, Biochemistry and*
560 *Pharmacology* 125:43–147.
- 561 Grzesiuk M, Mikulski A. 2006. The effect of salinity on freshwater crustaceans. *Polish*
562 *Journal of Ecology* 54(4):669–674.
- 563 Habig WH, Pabst MJ, Fleischner G, Gatmaitan, Z, Arias IM, Jakoby WB. 1974. The identity
564 of glutathione S-transferase B with ligandin, a major binding protein of liver. *Proceedings*
565 *of the National Academy of Sciences* 10:3879–3882.
- 566 Henry RP, Lucu C, Onken H, Weihrauch D. 2012. Multiple functions of the crustacean gill:
567 osmotic/ionic regulation, acid-base balance, ammonia excretion, and bioaccumulation of
568 toxic metals. *Frontiers in physiology* 3:431.
- 569 Hochachka PW, Fields J, Mustafa T. 1973. Animal life without oxygen: basic biochemical
570 mechanisms. *American Zoologist* 13(2):543–555.
- 571 Ivanina AV, Sokolov EP, Sokolova IM. 2010. Effects of cadmium on anaerobic energy
572 metabolism and mRNA expression during air exposure and recovery of an intertidal
573 mollusk *Crassostrea virginica*. *Aquatic toxicology* 99(3):330–342.

- 574 Ivanina AV, Taylor C, Sokolova, IM. 2009. Effects of elevated temperature and cadmium
575 exposure on stress protein response in eastern oysters *Crassostrea virginica*
576 (Gmelin). *Aquatic toxicology* 91:245–254.
- 577 Jacobs AFG, Jetten TH, Lucassen DC, Heusinkveld BG, Joost PN. 1997. Diurnal
578 temperature fluctuations in a natural shallow water body. *Agricultural and Forest*
579 *Meteorology* 88(1-4):269–277.
- 580 Jones DP. 2008. Radical-free biology of oxidative stress. *American Journal of Physiology-*
581 *Cell Physiology* 295(4):849–868.
- 582 Kalacheva GS, Gubanov VG, Gribovskaya IV, Gladchenko IA, Zinenko GK, Savitsky SV.
583 2002. Chemical analysis of Lake Shira water (1997–2000). *Aquatic ecology* 36(2):123-141.
- 584 Karaman GS, Pinkster S. 1977. Freshwater Gammarus species from Europe, North Africa
585 and adjacent regions of Asia (Crustacea-Amphipoda). Part I. Gammarus pulex-group and
586 related species. *Bijdr. Dierk* 47:1–97.
- 587 Kinne R, Schmitz JE, Kinne-Saffran E. 1971. The localization of the Na⁺ –K⁺-ATPase in
588 the cells of rat kidney cortex. *Pflügers Archiv* 329(3):191–206.
- 589 Laemmli U. Cleavage of structural proteins during the assembly of the head of
590 bacteriophage. 1970. *Nature* 227:680–685.
- 591 Larade K, Storey KB. 2002. A Profile of the Metabolic Responses to Anoxia in Marine.
592 *Sensing, Signaling and Cell Adaptation* 3:27.
- 593 Lesser MP. 2006. Oxidative stress in marine environments: biochemistry and physiological
594 ecology. *Annual Review of Physiology* 68:253–278.
- 595 Li E, Wang X, Chen K, Xu C, Qin J.G, Chen L. 2015. Physiological change and nutritional
596 requirement of Pacific white shrimp *Litopenaeus vannamei* at low salinity *Reviews in*
597 *Aquaculture* 7:1-19.
- 598 Livingstone DR. 1983. Invertebrate and vertebrate pathways of anaerobic metabolism:
599 evolutionary considerations. *Journal of the Geological Society* 140(1):27–37.

600 Matafonov DV. 2007. Ecology of Gammarus lacustris Sars (Crustacea: Amphipoda) in
601 Transbaikalian Water Bodies. *Biology Bulletin* 34(2):148–155.

602 Mates JM. 2000. Effects of antioxidant enzymes in the molecular control of reactive oxygen
603 species toxicology. *Toxicology* 153(1):83–104.

604 Morris S, van Aardt WJ, Ahern MD. 2005. The effect of lead on the metabolic and energetic
605 status of the Yabby, *Cherax destructor*, during environmental hypoxia. *Aquatic toxicology*
606 75:16–31.

607 Nelson DR, Fatland CL, Cardwell DL. 1977. Long-chain methylalkanes from haemolymph
608 of larvae of Japanese beetles, *Popillia japonica*. *Insect Biochemistry* 7(5-6):439–446.

609 O'Reilly CM, Sharma S, Gray DK, Hampton SE, Read JS, Rowley RJ, Schneider P, Lenters
610 JD, McIntyre PB, Kraemer BM, Weyhenmeyer GA, Straile D, Dong B, Adrian R, Allan
611 MG, Anneville O, Arvola L, Austin J, Bailey JL, Baron JS, Brookes JD, de Eyto E, Dokulil
612 MT, Hamilton DP, Havens K, Hetherington AL, Higgins SN, Hook S, Izmet'seva LR,
613 Joehnk KD, Kangur K, Kasprzak P, Kumagai M, Kuusisto E, Leshkevich G, Livingstone
614 DM, MacIntyre S, May L, Melack JM, Mueller-Navarra DC, Naumenko M, Noges P, Noges
615 T, North RP., Plisnier P-D, Rigosi A, Rimmer A, Rogora M, Rudstam LG, Rusak JA,
616 Salmaso N, Samal NR, Schindler DE, Schladow SG, Schmid M, Schmidt SR, Silow E,
617 Soyulu ME, Teubner K, Verburg P, Voutilainen A, Watkinson A, Williamson CE, Zhang
618 G. 2015. Rapid and highly variable warming of lake surface waters around the globe:
619 GLOBAL LAKE SURFACE WARMING. *Geophysical Research Letters* 42(10):773–781.

620 Philp A, Macdonald AL, Watt PW. 2005. Lactate—a signal coordinating cell and systemic
621 function. *Journal of Experimental Biology* 208(24):4561–4575.

622 Pörtner HO, Farrell AP. 2008. Physiology and climate change. *Science* 322(5902):690–692.

623 Pörtner HO, Hardewig I, Peck LS. 1999. Mitochondrial function and critical temperature in
624 the Antarctic bivalve, *Laternula elliptica*. *Comparative Biochemistry and Physiology Part*
625 *A: Molecular & Integrative Physiology* 124(2):179–189.

626 Pörtner HO, Heisler N, Grieshaber MK. 1984. Anaerobiosis and acid-base status in marine
627 invertebrates: a theoretical analysis of proton generation by anaerobic metabolism. *Journal*
628 *of Comparative Physiology B* 155(1):1–12.

629 Pörtner HO, Knust R. 2007. Climate change affects marine fishes through the oxygen
630 limitation of thermal tolerance. *Science* 315(5808):95-97.

631 Pörtner HO, Langenbuch M, Michaelidis B. 2005. Synergistic effects of temperature
632 extremes, hypoxia, and increases in CO₂ on marine animals: From Earth history to global
633 change. *Journal of Geophysical Research: Oceans* 110:9.

634 R Core Team. 2016. R: A language and environment for statistical computing. R
635 Foundation for Statistical Computing, Vienna, Austria. URL <https://www.R-project.org/>.

636 Rogozin DY, Genova SN, Gulati RD, Degermendzhy, AG. 2010. Some generalizations
637 based on stratification and vertical mixing in meromictic Lake Shira, Russia, in the period
638 2002–2009. *Aquatic ecology* 44(3):485-496.

639 Romano N, Zeng C. 2012. Osmoregulation in decapod crustaceans: implications to
640 aquaculture productivity, methods for potential improvement and interactions with elevated
641 ammonia exposure. *Aquaculture* 334:12–23.

642 Sars GO. 1863. *Om en i Sommeren 1862 foretagen zoologisk Reise i Christianias og*
643 *Trondhjems Stifter*. Dahl.

644 Schwarzenbach RP, Gschwend PM, Imboden DM. 2005. Environmental organic chemistry.
645 John Wiley & Sons.

646 Sokolova IM, Frederich M, Bagwe R, Lannig G, Sukhotin AA. 2012. Energy homeostasis
647 as an integrative tool for assessing limits of environmental stress tolerance in aquatic
648 invertebrates. *Marine Environmental Research* 79:1–15.

649 Sokolova IM, Frederich M, Bagwe R, Lannig G, Sukhotin AA. 2012. Energy homeostasis
650 as an integrative tool for assessing limits of environmental stress tolerance in aquatic
651 invertebrates. *Marine Environmental Research* 79:1-15.

652 Sokolova IM, Lannig G. 2008. Interactive effects of metal pollution and temperature on
653 metabolism in aquatic ectotherms: implications of global climate change. *Climate Research*
654 37(2-3):181–201.

655 Sokolova IM, Pörtner HO. 2003. Metabolic plasticity and critical temperatures for aerobic
656 scope in a eurythermal marine invertebrate (*Littorina saxatilis*, Gastropoda: Littorinidae)
657 from different latitudes. *Journal of Experimental Biology* 206(1):195–207.

658 Somero GN. 2004. Adaptation of enzymes to temperature: searching for basic
659 “strategies”. *Comparative Biochemistry and Physiology Part B: Biochemistry and*
660 *Molecular Biology*, 139(3):321–333.

661 Sutcliffe DW, Shaw J. 1967. The Sodium Balance Mechanism in the Fresh-Water
662 Amphipod, *Gammarus Lacustris* Sars. *Journal of Experimental Biology* 46(3):519–528.

663 Timofeyev M, Protopopova M, Pavlichenko V, Steinberg C. 2009. Can acclimation of
664 amphipods change their antioxidative response? *Aquatic ecology* 43:1041–1045.

665 Timofeyev M, Shatilina ZM, Bedulina DS, Protopopova, MV, Pavlichenko VV, Grabelnich
666 OI. 2008. Evaluation of biochemical responses in Palearctic and Lake Baikal endemic
667 amphipod species exposed to CdCl₂. *Ecotoxicology and environmental Safety* 70(1):99–
668 105.

669 Timofeyev MA. 2010. Ecological and physiological aspects of adaptation to environmental
670 factors in endemic Baikal and palearctic amphipods. Doctoral thesis, Tomsk Stat Univ.

671 Timoshkin OA, Suturin AN, Kravtsova LS, Kulikova NN, Malnik VV, Obolkina LA,
672 Shevelyova NG, Rozhkova NA, Bondarenko NA, Paradina LF, Milnik NG, Slugina ZV,
673 Proviz VI, Evstigneeva TD, Logacheva NF, Zaitseva EP, Nepokritich AV, Parfenova VV,
674 Kostornova TY, Boiko SM, Penzina MM, Semiturkina NA, Pobereshnaya AE, Basharina
675 TV, Potapskaya NV. 2008. Brief Results of International Investigations in the Berzovyi
676 Test Area (South Baikal) in the Period of 2000–2007 with Emphasis on Diversity,
677 Productivity of Benthos in the Shallow water Zone, and Their Controlling Factors, in
678 *Razvitie zhizni v protsesse abioticheskikh izmenenii na Zemle: Materialy nauchno-*
679 *prakticheskoi konferentsii (pos. Listvyanka Irkutskoi oblasti) (Life Development in*
680 *Response to Abiotic Changes on the Earth: Materials of the conference (Listvyanka, Irkutsk*
681 *Oblast). Novosibirsk: Rossijskaya Akademiya Nauk. Sibirskoe Otdelenie 344–357. (in*
682 *Russian)*

683 Timoshkin, OA, Sitnikova TY, Pronin NM, Proviz VI, Melnik NG, Kamaltynov RM,
684 Mazepova DF, Shoshnin AV. 2001. Index of animal species inhabiting Lake Baikal and its
685 catchment area. *Nauka, Novosibirsk* 1:74-113.

686 Väinölä R, Witt JDS, Grabowski M, Bradbury JH, Jazdzewski K, Sket B. 2007. Global
687 diversity of amphipods (Amphipoda; Crustacea) in freshwater. *Hydrobiologia* 595:241–
688 255.

689 Wang R, Zhuang P, Feng G, Zhang L, Huang X, Zhao F, Wang Y. 2013. The response of
690 digestive enzyme activity in the mature Chinese mitten crab, *Eriocheir sinensis* (Decapoda:
691 Brachyura), to gradual increase of salinity. *Scientia Marina* 77(2):323–329.

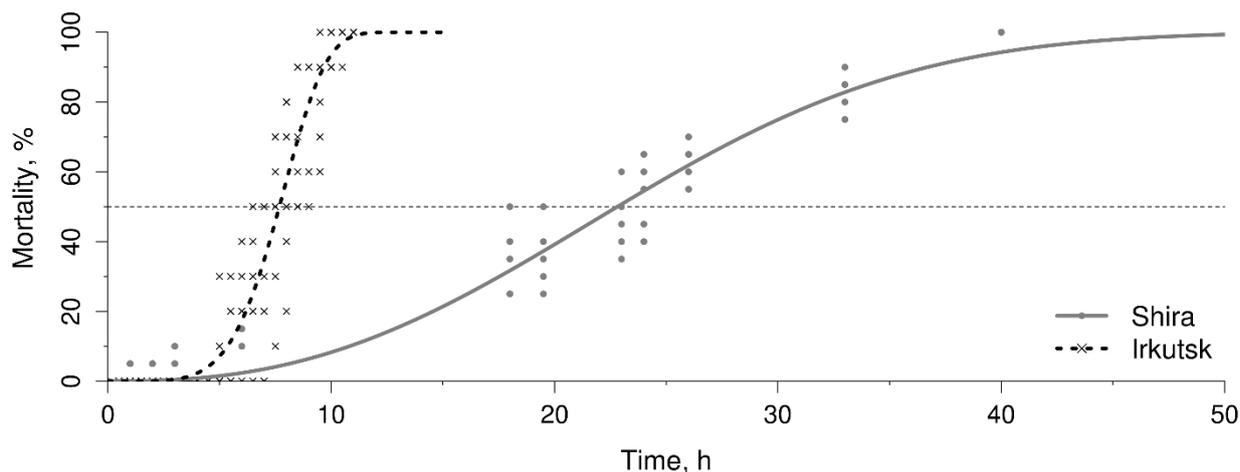
692 Wilhelm FM, Schindler DW. 1999. Effects of *Gammarus lacustris* (Crustacea: Amphipoda)
693 on plankton community structure in an alpine lake. *Can. Journal of Fisheries and Aquatic*
694 *Sciences* 56:1401–1408.

695 Wilson DL. 1994. The analysis of survival (mortality) data: fitting Gompertz, Weibull, and
696 logistic functions. *Mechanisms of ageing and development* 74(1):15-33.

697 Yemelyanova AY, Temerova TA, Degermendzhy AG. 2002. Distribution of *Gammarus*
698 *lacustris* Sars (Amphipoda, Gammaridae) in Lake Shira (Khakasia, Siberia) and laboratory
699 study of its growth characteristics. *Aquatic Ecology* 36(2):245–256.

700

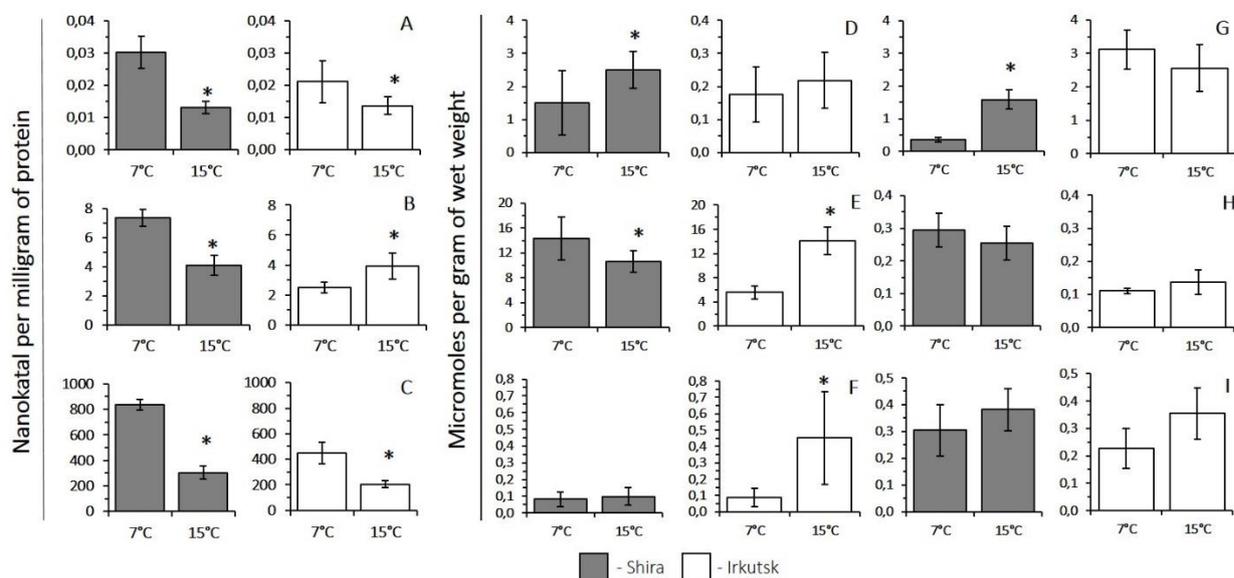
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703 **Fig.1.** Mortality rates in Lake Shira and Irkutsk Lake populations under heat shock

704 conditions. Data are fitted to the Weibull model.



705

706 **Fig.2.** The control levels in tissues of *G. lacustris* from saltwater (Lake Shira) and freshwater

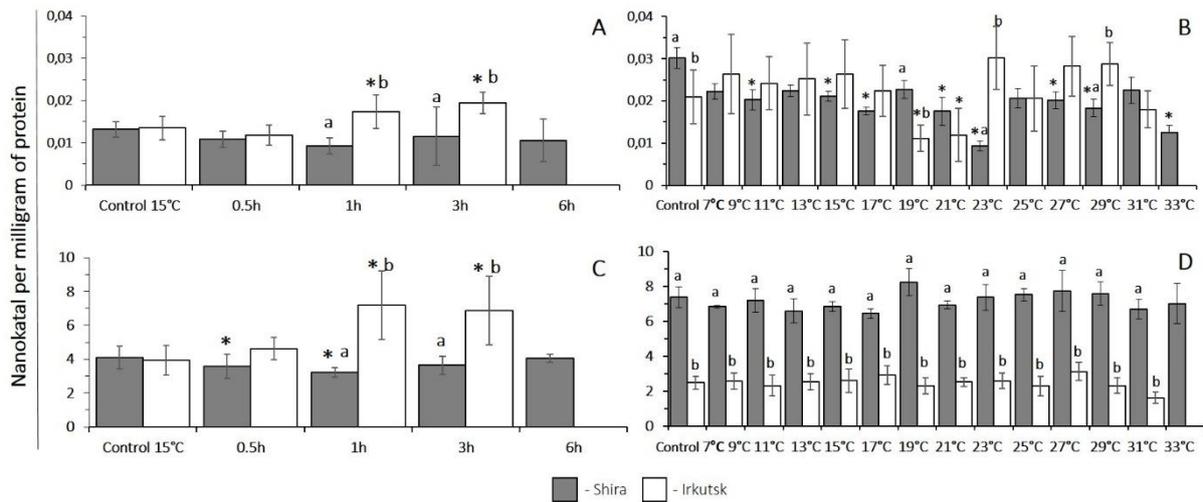
707 (a lake in Irkutsk) populations at 7°C and 15°C. A – peroxidase, B – glutathione S-

708 transferase, C – catalase, D – glucose, E – glycogen, F – lactate, G – ATP, H – AMP, I –

709 ADP. Different letters above the columns indicate differences between populations at a given

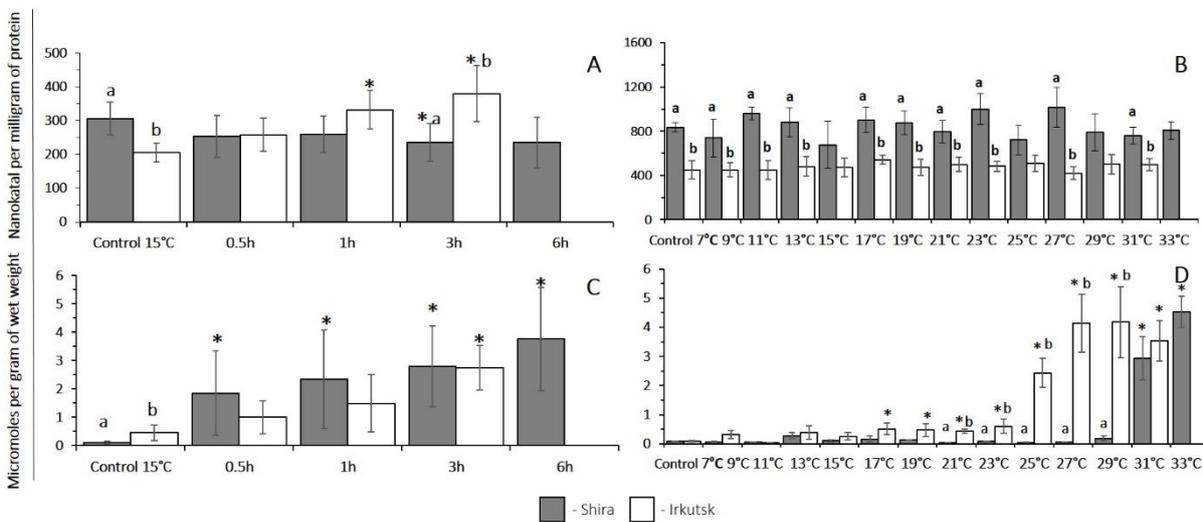
710 time point ($P < 0.05$).

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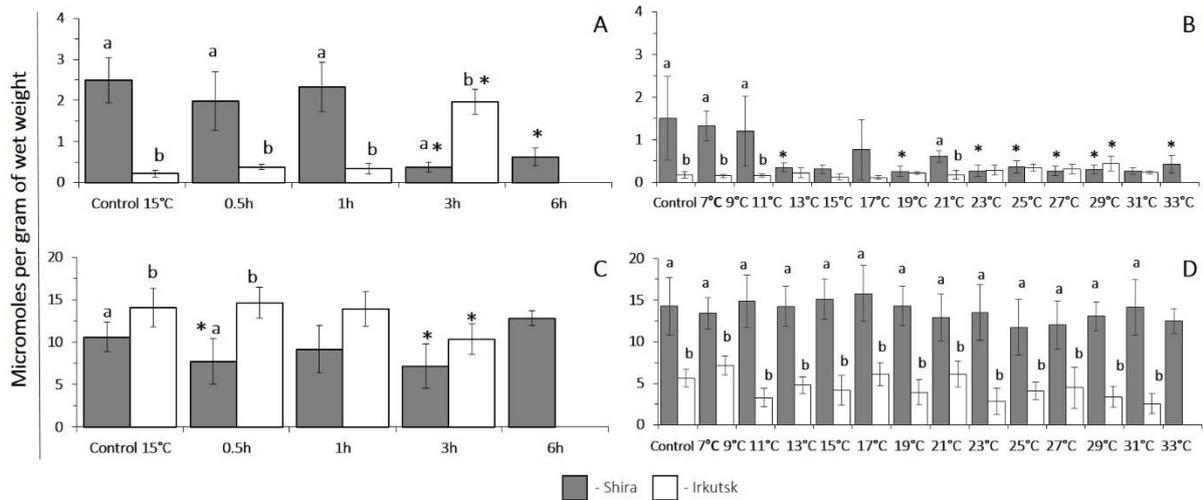
713 **Fig.3.** Peroxidase and glutathione S-transferase activity levels in *G. lacustris* from saltwater
 714 (Lake Shira) and freshwater (a lake in Irkutsk) populations. A, B – peroxidase activity, C, D -
 715 glutathione S-transferase activity. A, C - acute exposure to 30°C (control 15°C). B, D – gradual
 716 warming (1°C h⁻¹) (control 7°C). Different letters above the columns indicate differences
 717 between populations at a given time point. Asterisks (*) denote a significant change during the
 718 thermal exposure compared to the control animals from the same population (P<0.05).



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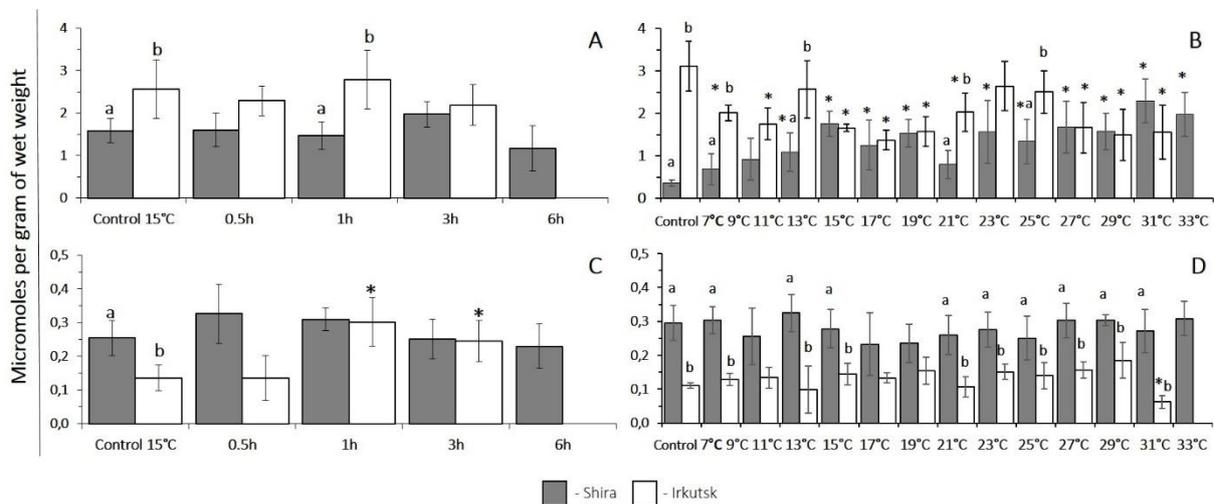
721 **Fig.4.** Catalase activities and lactate content in *G. lacustris* from saltwater (Lake Shira) and
 722 freshwater (a lake in Irkutsk) populations. A, B – catalase activity, C, D – lactate content. A, C
 723 - acute exposure to 30°C (control 15°C). B, D – gradual warming (1°C h⁻¹) (control 7°C).
 724 Different letters above the columns indicate differences between populations at a given time
 725 point. Asterisks (*) denote a significant change during the thermal exposure compared to the
 726 control animals from the same population (P<0.05).



727

728 **Fig.5.** Glucose and glycogen content in amphipods *G. lacustris* from saltwater (Lake Shira) and
 729 freshwater (a lake in Irkutsk) populations. A, B – glucose, C, D – glycogen. A, C - acute
 730 exposure to 30°C (control 15°C). B, D – gradual warming (1°C h⁻¹) (control 7°C). Different
 731 letters above the columns indicate differences between populations at a given time point.
 732 Asterisks (*) denote a significant change during the thermal exposure compared to the control
 733 animals from the same population (P<0.05).

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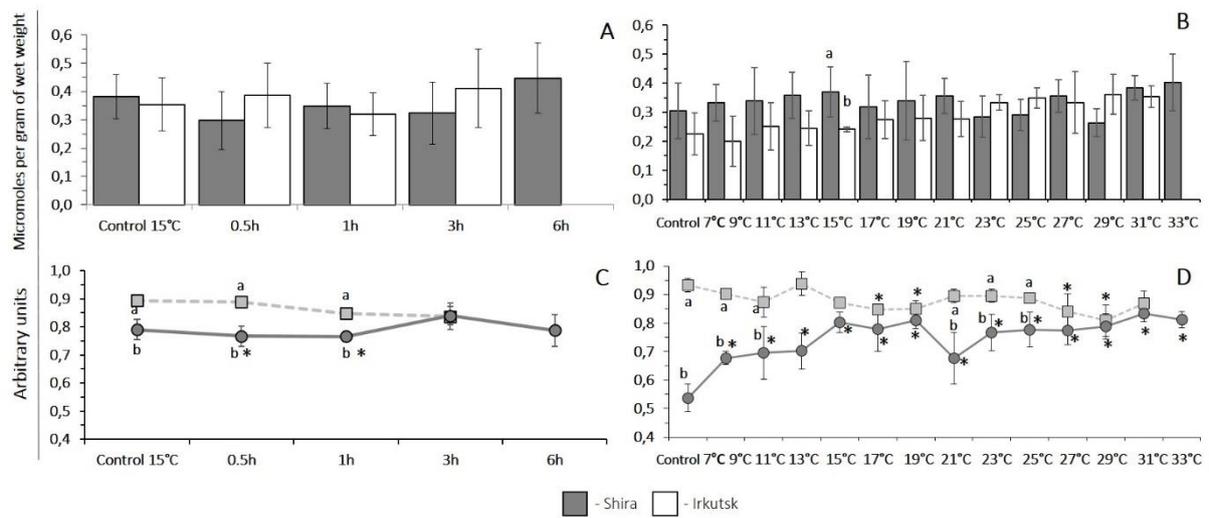


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736

737 **Fig.6.** ATP and AMP levels in tissues of *G. lacustris* from saltwater (Lake Shira) and freshwater
 738 (a lake in Irkutsk) populations. A, B – ATP, C, D – AMP. A, C - acute exposure to 30°C
 739 (control 15°C). B, D – gradual warming (1°C h⁻¹) (control 7°C). Different letters above the
 740 columns indicate differences between populations at a given time point. Asterisks (*) denote a

741 significant change during the thermal exposure compared to the control animals from the same
 742 population ($P < 0.05$).



743

744 **Fig. 7.** ADP and Adenylate energy charge (AEC) levels in tissues of *G. lacustris* from saltwater
 745 (Lake Shira) and freshwater (a lake in Irkutsk) populations. A, B – ADP, C, D – AEC. A, C -
 746 acute exposure to 30°C (control 15°C). B, D – gradual warming (1°C h⁻¹) (control 7°C).
 747 Different letters above the columns indicate differences between populations at a given time
 748 point. Asterisks (*) denote a significant change during the thermal exposure compared to the
 749 control animals from the same population ($P < 0.05$).